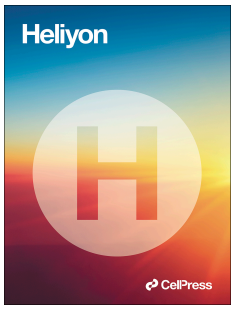

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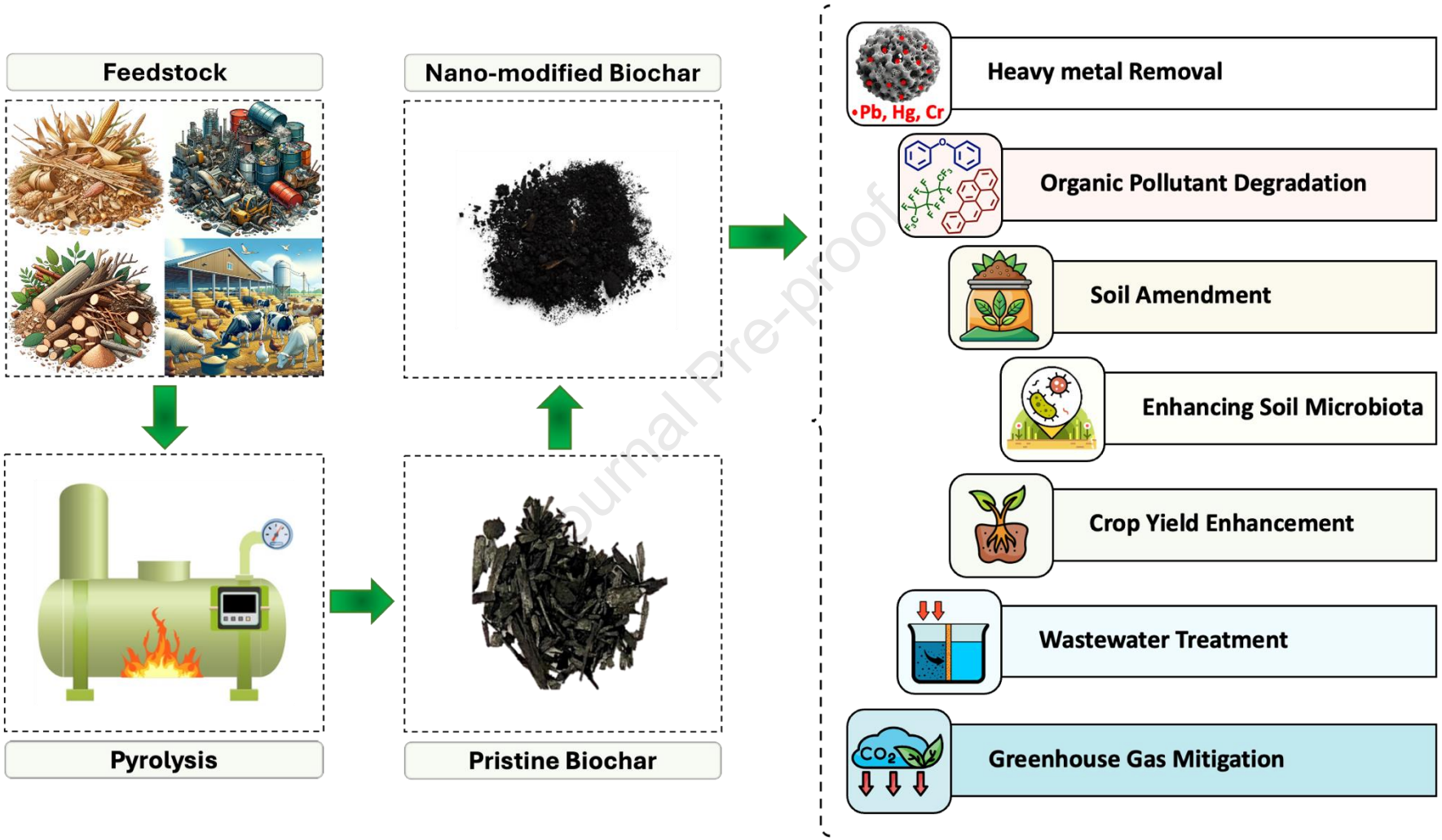

Exploring Nanomaterial-modified Biochar for Environmental Remediation Applications

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Abstract

Environmental pollution, particularly from heavy metals and toxic elements, poses a significant threat to both human health and ecological systems. While various remediation technologies exist, there is an urgent need for cost-effective and sustainable solutions. Biochar, a carbon-rich product derived from the pyrolysis of organic matters, has emerged as a promising material for environmental remediation. However, its pristine form has limitations, such as low adsorption capacities, a relatively narrow range of pH adaptability which can limit its effectiveness in diverse environmental conditions, and a tendency to lose adsorption capacity rapidly in the presence of competing ions or organic matters. This review aims to explore the burgeoning field of nanomaterial-modified biochar, which seeks to overcome the limitations of pristine biochar. By incorporating nanomaterials, the adsorptive and reactive properties of biochar can be significantly enhanced. Such modifications, especially biochar supported with metal nanoparticles (biochar-MNPs), have shown promise in various applications, including the removal of heavy metals, organic contaminants, and other inorganic pollutants from aqueous environments, soil, and

air. This review provides a comprehensive overview of the synthesis techniques, characterization methods, and applications of biochar-MNPs, as well as discusses their underlying mechanisms for contaminant removal. It also offers insights into the advantages and challenges of using nanomaterial-modified biochar for environmental remediation and suggest directions for future research.

Keywords: Nanomaterial-modified biochar, biomass, metal nanoparticles, environmental remediation, nanotechnology

1. Introduction

In recent years, there has been growing concerns over the escalating contamination of soil, water, and air. One of the most pressing concerns is the contamination of soil and water with heavy metals and toxic elements, such as lead (Pd), arsenic (As), mercury (Hg), and chromium (Cr), in natural ecosystems [1-7]. These elements are not only highly toxic but also resistant to degradation, leading to their accumulation in environment. The bioaccumulation of these toxicants can have detrimental effects on human health, often through the soil-plant-animal-human biological chain [8]. Air pollution is another critical environmental concern. Driven by industrial activities, vehicle emissions, and agricultural practices, air pollution introduces a variety of harmful substances into the atmosphere, including nitrogen oxides (NO_x), sulfur oxides (SO_x), volatile organic compounds (VOCs), and greenhouse gases (GHGs) such as carbon dioxide (CO_2) and methane (CH_4) [9]. These pollutants can cause severe health problems, including respiratory and cardiovascular diseases, and contribute to global climate change [10, 11]. The persistence and bio-accumulative nature of these contaminants make it imperative to develop effective and sustainable remediation technologies. For example, remediation of heavy metal-contaminated soil and water has primarily relied on conventional techniques such as chemical precipitation, ion exchange, and membrane

filtration [12]. While these methods have shown some efficacy, they come with their own challenges, such as use of additional reagents and high membrane costs and membrane fouling. Micro/nano plastic pollutions further highlight the need for effective remediation strategies to mitigate environmental and health risks [13].

In the context of environmental pollution, biochar has emerged as a promising solution to various environmental challenges [14]. Biochar is rich in carbon, fine-grained, porous and has large surface areas. It is produced by pyrolyzing biomass materials, such as plant matters, industrial byproducts, agricultural waste, sewage sludge, and animal manure, in an environment devoid of oxygen or with limited oxygen, at temperatures not exceeding 1000 °C, as shown in **Figure 1** [15-17]. Traditionally used in agriculture as soil amendments, biochar has gained attention for its multifaceted applications especially for environmental remediation [18, 19]. One of the key advantages of biochar is its high adsorption capacity, owing to its large specific surface area and porous structure. These properties make biochar particularly effective for the adsorption of heavy metals and toxic elements from contaminated soil and water [20]. Biochar is sourced from a variety of renewable biomass wastes, making it a sustainable and low-cost alternative to other remediation materials [21, 22]. However, the use of biochar has several challenges. While it offers a cost-effective and environmentally friendly solution, its efficiency in remediating high-concentration contaminants has yet to be satisfying. Moreover, the application of large amounts of biochar can sometimes lead to soil nutrient imbalances, requiring careful consideration of its use in different environmental settings. Pristine biochar has its own limitations, one of which is its relatively low adsorption capability for heavy metals and toxic elements. This has led to the exploration of various modification strategies aiming at enhancing its adsorptive properties. These include changing surface charges of biochar using metal nanoparticles, expanding its specific surface areas

through activating agents, and incorporating metal (or metal oxides) that can interact with pollutants [23-25].

The advent of nanotechnology and the use of nanomaterials have opened new avenues to enhance the effectiveness of pristine biochar. Incorporating metal nanoparticles (MNPs) into biochar, resulting in biochar-MNPs , is a promising approach to improve its ability to absorb and remove contaminants [26, 27]. Nanostructured materials consist of a broad category of substances that have at least one dimension measuring at the nanometer scale. These materials possess structural characteristics that fall somewhere between those of bulk solids and molecules. As a result, they display a unique set of physical, optical, and chemical properties that are distinct from those of their larger or macroscopic counterparts [28, 29]. By incorporating nanoparticles into the biochar matrix, researchers have been able to create a material with synergistic properties that far exceed those of pristine biochar. The nanoparticles serve multiple functions in this hybrid material. They can increase the surface area and porosity of the biochar, thereby enhancing its adsorption capacity. Additionally, nanoparticles can introduce new functional groups to the biochar, increasing its chemical reactivity, and enabling more effective binding with heavy metals and toxic elements. These nanoparticles, due to their large surface-to-volume ratios compared to the bulk materials, are highly effective against a wide range of pollutants, especially heavy metals and organic compounds [30]. However, they tend to clump together into larger particles, reducing their effectiveness [31]. To address this, biochar has emerged as a popular supporting material for these nanoparticles, given its high surface area and ion-exchange capacities [32, 33]. Combining biochar with metal nanoparticles creates a novel composite which maximizes the benefits of both materials while minimizing their drawbacks. Such composites not only stabilize the nanoparticles but also enhance their functional properties. Additionally, nanomaterial-modified biochar has the potential

to improve air quality by reducing emissions of harmful gases, such as NOx and VOCs, through enhanced adsorption and catalytic processes. Furthermore, it can contribute to mitigating greenhouse gases by sequestering CO_2 and reducing CH_4 emissions [34]. Biochar's ability to adsorb air pollutants is influenced by its physiochemical properties, such as surface area and functional groups, which can be enhanced through various activation processes and pyrolysis conditions.

This review aims to cover the synthesis techniques for biochar-assisted nanomaterials and their applications in removal of various contaminants from the environment. We also discuss the mechanisms behind its effectiveness and suggest directions for future research. Additionally, we explore the potential of these materials to enhance air quality and reduce greenhouse gas emissions, offering insights into their multifaceted role in environmental remediation and sustainable practices.

2. Nano-modified Biochar Preparation Techniques

Extensive research has been conducted to couple the benefits of biochar's porous structure with the unique physical attributes of nanomaterials, thereby enhancing its capacity to remove both organic and inorganic contaminants. Two primary approaches exist for creating nano-modified biochar with advanced functionalities as depicted in **Figure 2**: one involves modifying the initial biomass before its conversion into biochar through processes like pyrolysis, calcination, or coprecipitation; the other one directly alters the already prepared biochar [31, 35-37]. These modifications often include surface treatments using various metal precursors, nanoparticles, and organic-inorganic polymers, resulting in biochar nanohybrids. The properties of these biochar-MNPs are influenced by type of feedstock, pyrolysis temperature, ratio of biochar to metal

115 nanoparticles, and pyrolytic reaction media. The modifications often aim to maximize biochar
116 yields by optimizing temperature, heating rate, and residence time based on specific feedstock
117 characteristics [38, 39]. The specific surface area (SSA) values for most nano-modified biochars
118 range significantly, from 5.58 to 1736 m^2/g, attributed to their small particle sizes. Research
119 indicates that nano-modified biochar's SSA can rise with an increase in pyrolysis temperature.
120 However, there are different views, with some suggesting that an increase in pyrolysis temperature
121 does not impact or could even decrease the SSA of nano-modified biochars. These discrepancies
122 could stem from variations in the original materials used to produce the nano-modified biochars
123 [40]. Enhancing biochar/biochar-MNPs stability involves post-pyrolysis treatments such as
124 chemical regeneration or physical modification, which not only increase its resistance to
125 degradation but also its efficacy in environmental applications. Chemical regeneration is a key
126 method for reactivating biochar, especially for adsorbing heavy metals and organic pollutants like
127 dyes and antibiotics. The process typically employs solvents and chemical reagents to desorb these
128 pollutants. The efficiency of regeneration depends on the type of pollutants, and different solvents
129 are used for organic and inorganic adsorbates. For instance, in a study involving the removal of
130 lead from wastewater, biochar was regenerated using solutions like sodium hydroxide and
131 hydrochloric acid. The removal efficiency varied across different cycles, generally showing a
132 slight decrease in efficiency over successive cycles. This method ensures biochar maintains its
133 adsorption activity, making it a sustainable option for wastewater treatment [41].
134

135 **2.1 Biomass Modification**

136 Biomass pretreatment entails altering raw feedstock before it is converted into biochar
137 through processes like pyrolysis and calcination. This approach is deemed energy-efficient

because it allows for the concurrent processing of both the biomass and metal-containing precursors. During such alteration processes, the original biomass may be subjected to surface modifications using a range of metal-containing compounds, specifically metal oxides, metal chlorides, metal sulfides, metal sulfates, and metal nitrates, along with nanoparticles and organic-inorganic polymers. These modifications facilitate the creation of biochar-supported nanomaterials, details of which are elaborated in the following section.

Using metal salts to pretreat biomass prior to pyrolysis is a technique for generating biochar-MNPs. Biomass can be impregnated with various metal salts like $FeCl_3$, $Fe(NO_3)_3$, $AlCl_3$, $MgCl_2$, $MnCl_2$, $CaCl_2$, and $ZnCl_2$, as well as sulfur-based and organic-based metal salt solutions, either with or without the application of an electric field [42-46]. This allows metal ions in the solutions to deposit onto the surface or interior of the feedstocks. Following this, the pretreated biomass undergoes pyrolysis at temperatures ranging from 400 to 800 °C in an environment with limited or no oxygen. During the pyrolysis process, these metal ions transform into metal oxide nanoparticles like Fe_2O_3, Al_2O_3, MgO, MnO_2, CaO, and ZnO, or into zero-valent metals that adhere to the biochar surface. For example, in a study reported by Chaukura et al. [47], Fe_2O_3-biochar nano-composite was prepared by pyrolyzing $FeCl_3$-impregnated paper and paper sludge (PPS) at 750 °C. The PPS biomass was immersed in a $FeCl_3$ solution for 2 h and then air-dried at 80 °C for another 2 h. The $FeCl_3$-impregnated PPS, with a mass-to-volume ratio of 1:3, was pyrolyzed at 750 °C for 2 h in a N_2 atmosphere, yielding Fe_2O_3-biochar. This nanocomposite was used for the efficient removal of methyl orange (MO) from contaminated wastewater and the results showed that its MO adsorption capacity was 52.79 % higher than that of pristine biochar. The surface area and porosity measured by the Brunauer–Emmett–Teller (BET) method were lower for the Fe_2O_3-biochar composite compared to the pristine biochar. This decrease could be

linked to the clogging of the biochar's pores by Fe_2O_3 and other inherent metallic compounds in the absorbent material. Similarly, Fe_3O_4-based magnetic biochar was synthesized by pyrolyzing $FeCl_3$-treated corn straw under an electric field at 600 °C for 1 h. This process resulted in the formation of uniformly dispersed, rod-like Fe_3O_4 nanoparticles on the biochar's surface. Notably, the Fe_3O_4-based magnetic biochar demonstrated a remarkable capacity (113 mg/g) for adsorbing lead [48]. Bismuth-impregnated biochar was produced through the thermal pyrolysis of bismuth oxide and hydrochloric acid-treated wheat straw, at temperatures ranging from 400 to 600 °C [49]. This process, conducted at a heating rate of 10 °C/min in a N_2 atmosphere for 1 h, resulted in biochar with a unique porous structure, featuring pores ranging from 0.5 to 1 μm in diameter. This bismuth-modified biochar was noted for its high capacity in adsorbing phosphorus (125.40 mg/g), arsenic (16.21 mg/g), and chromium (12.23 mg/g). This study investigated the use of bamboo biochar, functionalized with Mg-Al and Mg-Fe layered double hydroxides (LDHs), for the efficient removal of phosphate from aqueous solutions. The composite containing 40% Mg-Al LDH exhibited the highest phosphate adsorption, achieving over 95% saturation within 1 h. The research highlighted the potential of using such biochar composites not only for effective wastewater treatment but also as a recyclable material for agricultural applications, promoting sustainable environmental management. In another study reported by Omidvar-Hosseini et al., $Ni_{0.5}Zn_{0.5}Fe_2O_4$ magnetic nanoparticles supported on Acacia Nilotica seed shell ash (ANSA) were synthesized and applied for effective adsorption of Pb(II) ions from water [50]. An aqueous mixture of metal nitrates $Fe(NO_3)_3 \cdot 9H_2O$, $Zn(NO_3)_2 \cdot 6H_2O$, $Ni(NO_3)_2 \cdot 6H_2O$, and ANSA was dissolved in DI water, to which extracted egg-white solution was added. Following 30 min of continuous stirring, the sol-gel obtained was dried at 80 °C to yield dry precursors. These precursors were then finely ground and calcined at 550 °C for 2 h in a muffle furnace. Various

characterization techniques like Fourier Transform Infrared Spectrometry (FTIR), Scanning Electron Microscopy (SEM), X-ray Diffraction (XRD), and Vibrating Sample Magnetometer (VSM) confirmed the successful incorporation of nanoparticles into ANSA, forming less than 100 nm particles with superparamagnetic behavior. The Langmuir model best described the adsorption process, and the kinetics fit the pseudo-second-order rate equation. The nanoparticles could be easily separated from water using an external magnet, making them a practical and eco-friendly solution for treating water pollution caused by Pb (II) ions. Similarly, Li group synthesized nano ZnO/ZnS-modified biochar via slow pyrolysis of zinc-contaminated corn stover biomass [46]. Slow pyrolysis was conducted in a lab-scale fixed-bed reactor, using zinc-loaded biomass in a horizontal tubular reactor (60 mm I.D., 300 mm length). The reactor was first purged with N_2 gas to eliminate O_2. Biochar from raw corn stover was then produced at four different temperatures: 500, 600, 700, and 800 °C. This modified biochar, uniformly coated with nano ZnO/ZnS, displayed superior porosity (with a BET surface area of 397.4 m^2/g and a total pore volume of 0.43 cm^3/g) when compared to the unmodified biochar (with a BET surface area of 102.9 m^2/g and a total pore volume of 0.20 cm^3/g). Batch absorption tests indicate that the obtained biochar modified with nano ZnO/ZnS exhibited a robust ability to absorb Pb (II), Cu (II), and Cr (VI), with maximum absorption capacities that were notably greater than those of common biochar. Biochar impregnated with magnesium was prepared by pyrolyzing biomass feedstock (cow dung) and utilized to reduce the leaching of phosphorus from the soil [51]. The cow dung biomass mixed with $MgCl_2 \cdot 6H_2O$ solution was placed in porcelain crucibles, which were then sealed and pyrolyzed at 600 °C for 1 h in an O_2-limited muffle furnace. This process was aimed at producing Mg-modified biochar. The data revealed that nanoscale magnesium oxides were uniformly distributed on the biochar's surface. The MgO-biochar composite had a mesoporous structure with

an average pore size of 1.74 μm and could effectively minimize the loss of phosphorus through leaching from the soil and offer improved stability.

Using a low-temperature method, coprecipitation of biomass and metal salts can serve as another technique for modifying char. A study by Gupta and colleagues introduced a feasible way to alter the surface of orange peel powder (OPP) by coprecipitating it with ferric chloride hexahydrate and ferrous sulfate heptahydrate. This led to the creation of a novel magnetic nano-adsorbent Fe_3O_4−OPP (MNP−OPP), which showed the covalent binding between the hydroxyl groups in MNP and the carboxyl groups in OPP. The nanocomposite was further used for cadmium ion removal from aqueous solutions [52]. It showed that under optimal conditions MNP-OPP achieved a maximum removal of Cd^{2+} at 76.92 mg/g based on Langmuir Model and the adsorption was thermodynamically favorable.

Another approach for creating biochar-MNPs involved directly subjecting biomass rich in target metal elements to pyrolysis. This process transformed the metal elements in the biomass into metal nanoparticles. For instance, biochar rich in calcium (CRB) was prepared by carbonizing crab shells at a temperature ranging from 300 to 900°C at a heating rate of 10 °C/min for 2 h [53]. Elemental analysis showed a high calcium content, but a low carbon content in the crab shell biochar. TG indicated that calcite-based CRB was formed at 600 °C, while lime-based CRB was generated at 700 °C. The nanocomposites showed the phosphorus adsorption efficiency varied from 26% to 100% for a phosphate solution (80 mg/L) and from 11% to 63% for anaerobic digestion effluent. These findings suggest that CRB was effective for removing or recovering phosphorus from wastewater.

In addition to the use of metal oxides/salts, biochar coated with multi-walled carbon nanotubes (CNT) were was also reported [54]. Biochar-CNT was produced through slow pyrolysis

of carboxyl-functionalized CNT-impregnated feedstocks, including milled hickory chips and sugarcane bagasse. These CNT-treated feedstocks were placed in a quartz tube within a tubular furnace (MTI, Richmond, CA) and subjected to pyrolysis at 600 °C for 1 h, with a heating rate of 10 °C/min, in a N_2-rich environment. It was found that CNTs improved the biochar's physicochemical attributes, such as surface area, porosity, and thermal stability. Batch sorption test on methylene blue showed efficient adsorption.

Recent studies have increasingly explored the calcination of biomass for biochar and composite material synthesis. This process involves heating inorganic materials to high temperatures to improve crystallinity and remove surface impurities and volatiles. An example of this is the successful synthesis of the sugarcane bagasse-Fe_3O_4 composite through coprecipitation of bagasse with $FeSO_4 \cdot 7H_2O$ and $FeCl_3 \cdot 6H_2O$ in the presence of ammonia, followed by calcination [55]. TEM identified spherical or rod-shaped structures in the altered bagasse. A shift in the IR spectrum peak from 897 cm^{-1} to 874.6 cm^{-1} verified the interaction between bagasse and Fe-OH, while the appearance of Fe-O vibration bands confirmed the composite's formation. Furthermore, absorption bands at 712.2 cm^{-1}, 585.5 cm^{-1}, and 436.9 cm^{-1} were attributed to Fe-O vibrations, a consequence of high-temperature calcination.

Table 1 summarizes the various biomass modifications discussed in the section, providing a comprehensive overview of the feedstocks used, the types of metal precursors involved, the conditions under which pyrolysis or calcination was carried out, the specific nanocomposites formed, and their respective applications. This summary table serves as a quick reference for understanding the diversity of approaches in biomass modification for environmental remediation

and pollution control, highlighting the efficiency of each method in creating effective nanocomposites for targeted applications.

2.2 Biochar Modification

2.2.1 Modification Using Metal Precursors

The common approach to acquire biochar-MNPs is enhancing the surface chemistry of biochar through modification with metal precursors. Initially, biomass undergoes transformation into biochar through processes like pyrolysis, hydrothermal carbonization, and gasification. Following this, the biochar is immersed in solutions of metal salts, allowing metal ions to adhere to the biochar's pores and surface. Ultimately, through processes such as coprecipitation, pyrolysis, and pH adjustment, metal nanoparticles are deposited onto the biochar's surface, creating biochar-MNPs. The incorporation of metallic entities into the carbon matrix of biochar not only enhances its surface properties but also facilitates the easier recovery of the adsorbent material after use. For instance, manganese dioxide nanoparticles were loaded onto biochar derived from pyrolyzed invasive water hyacinth using a redox precipitation method, yielding MnO_2-biochar effective for the adsorption of Cd (II), Cu (II), Zn (II), and Pb (II) [56]. Compared to biochar alone that showed negligible adsorption of heavy metal ions, the MnO_2-biochar nanocomposite demonstrated substantial capacity for adsorbing these ions. Similarly, the coprecipitation process involving the ethanol-mediated reduction of $KMnO_4$ in a biochar suspension led to the creation of the MnO_2-biochar nanocomposite with enhanced porosity [57]. The pore volume of this nanocomposite was approximately three times smaller than that of biochar and 1.5 times smaller than that of nano MnO_2. In a study, sewage sludge-derived biochar was first produced by pyrolyzing the sludge at 600 °C for 3 h a 10 °C/min heating rate in a N_2 atmosphere. Subsequently, zero-valent iron

nanoparticles were immobilized to the biochar, creating SSB-nZVI-biochar [58]. This composite showed remarkable efficiency in removing Cr^{6+} and Pb^{2+} ions from solutions, achieving approximately 90% and 82% removal within 30 min for Cr^{6+} and Pb^{2+}, respectively. Kim et al. modified the Miscanthus biochar with amorphous iron (hydr)oxide to remove arsenite from water [59]. Miscanthus was initially pyrolyzed at 500 °C for 1 h, using a 10 °C/min heating rate in a N_2 atmosphere to produce biochar. Then, to prepare Fe-modified biochar, Fe ions were precipitated onto the biochar by adjusting the pH to 9 using NaOH. This Fe-modified biochar demonstrated enhanced effectiveness in adsorbing arsenite (56.06 and 47.90 mg/g Fe for powder and beads, respectively). In their research, Wan et al. utilized bamboo biochar, which was pyrolyzed at 600 °C for 2 h subsequently functionalized with Mg-Al and Mg-Fe layered double hydroxides (LDHs), to effectively remove phosphate from water solutions [60]. The composite containing 40% Mg-Al LDH exhibited the highest phosphate adsorption, achieving over 95% saturation within 1 h. Tomczyk et al. modified sunflower husk biochar with AgNPs for efficient tetracycline removal [61]. In another study, they effectively created a new electrochemical sensor for the detection of nitrite, achieved by immobilizing gold nanoparticles to biochar [62]. The resulting biochar/Au composite displayed distinctive characteristics, including a vast surface area, numerous surface functional groups, and outstanding conductivity, all contributing to the enhanced performance of the modified fluorine-doped Titanium oxide electrode in sensing applications. Wang et al. developed a biochar derived from the seedling of white myoga ginger decorated with gold nanoparticles (W-biochar/Au) [63]. The as-prepared W-biochar was mixed into the AuNPs colloidal solution and then subjected to pyrolysis again at a specific temperature (700, 850, or 950 °C) for 2 h. This process resulted in the formation of the W-biochar/Au nanocomposite. The nanocomposite formed at 850 °C, was employed for the detection of hydroquinone and catechol,

demonstrating extremely low detection thresholds and enhanced sensitivities. This nanocomposite displayed a highly porous configuration, excellent water dispersibility, and an extensive specific surface area. The notable performance of the W- biochar/Au nanocomposite is likely attributed to its pronounced porous architecture and increased relative contents of C-O/N-O, C=O, and COOH/COOR bonds. Zhu et al. developed biochar decorated with magnetic nanoscale zero-valent iron (nZVI-biochars), featuring varying Fe/biochar impregnation ratios [64]. By systematically analyzing the mechanisms involved in the removal of heavy metals, they conclusively demonstrated that nZVI-biochar had superior removal efficiency for various heavy metals (Pb^{2+}, Cd^{2+}, Cr^{6+}, Cu^{2+}, Ni^{2+}, Zn^{2+}). The commendable magnetic properties of the developed material facilitate the easy recycling of nZVI-biochar from wastewater. In a study by Liao et al., pineapple peel waste-based biochar is decorated with both magnetic Fe_2O_3 and lanthanum hydroxide ($La(OH)_3$) and used as phosphate absorbent [65]. As $La(OH)_3$ loading in the composite was increased, the adsorption capacity for phosphate was increased up to 101.16 mg/g although its magnetic property was reduced. The $La(OH)_3$-modified magnetic biochar had adsorption capacity 27 times greater than the pristine biochar and substantially surpassed most phosphate adsorbents in capacity. **Table 2** presents a comprehensive overview of modified biochar nanocomposites, delineating their feedstock sources, metal precursors used for modification, the pyrolysis techniques employed, and their subsequent applications for environmental remediation.

2.2.2 Other Modification Techniques

Another method of fabrication of biochar involves impregnating biochar with various organic or inorganic polymers. This impregnation transforms the biochar into an effective composite, enhancing its porosity and interface chemistry with heavy metal pollutants. For instance, a bionanocomposite, consisting of an organic-inorganic blend of chitosan, nanoclay, and

bark chip-derived biochar, was selected for the simultaneous immobilization of Cu, Pb, and Zn metal ions in contaminated soil and water environments [66]. This bionanocomposite was prepared by adding bark chip-derived biochar to the mixture of nanoclay/chitosan solution. The composite showed a mixed exfoliated/intercalated morphology with the incorporation of minimal nanoclay amounts (5 wt%). The batch adsorption tests revealed that the nanocomposite's adsorption capabilities for Cu^{2+}, Pb^{2+}, and Zn^{2+} significantly surpassed those of the unmodified biochar. As per the FTIR analysis, the predominant mechanism for metal immobilization is the interaction with $–NH_2$ groups. Chitosan was utilized as an organic polymer due to its potential biocompatibility, environmental friendliness, and its appropriateness for biocomposite preparation [67]. Nanoclay, with its exchangeable hydrated cations in a layered structure, is commonly applied as an additive to enhance various physical and adsorptive properties of polymers like chitosan [68].

During the preparation of bulk biochar, nanobiochar inherently forms, though its yield is minimal (e.g., ~2.0% in peanut shell biochar).[69] To enhance the nanoparticle content in biochar, a process for size reduction is essential, which can be conducted effortlessly through grinding or milling. Prior studies have employed hand grinding with mortars to decrease the size of various bulk biochars.[70] Ball milling has been underscored as a promising method to boost the yield and reduce the particle size of nano-modified biochar for large-scale production.[71, 72] Several key ball milling parameters are ball diameters, mass ratios of balls to biochar, milling speed, duration of milling, temperature, and liquid media.[73]

While nano-modified biochars have demonstrated improved adsorption properties, it is anticipated that the incorporation of a multi-cationic salt can further enhance adsorption by introducing a dual adsorption mechanism to the biochar. For example, in a study g-C_3N_4/$FeVO_4$ (CI) and g-C_3N_4/$FeVO_4$/Fe@NH_2-biochar (CIB) nano-hetero assemblies were synthesized for the

342 removal of methyl paraben (MeP) and 2-chlorophenol (2-CP) using a combination of adsorption
343 photocatalysis and photo-ozonation [74]. The study demonstrated excellent results, achieving
344 98.4% degradation of MeP and 90.7% degradation of 2-CP under simultaneous adsorption and
345 photocatalysis in the presence of CIB. In their research, Wu et al. developed biochar-supported
346 Ag/Fe nanoparticles (Ag/Fe/MB) for the removal of cephalexin (CLX) in aqueous solution. The
347 process involved synthesizing nano zerovalent iron (nZVI) with biochar using $FeSO_4 \cdot 7H_2O$ and
348 $NaBH_4$. This was then treated with $AgNO_3$ under vigorous stirring, leading to the formation of a
349 discontinuous Ag layer on the nZVI/ biochar surface. The Ag/Fe/MB composite was highly
350 effective in removing CLX from aqueous solution. Specifically, it was found that more than 86%
351 of CLX was removed by Ag/Fe/MB in 90 min under the conditions of 1.5 g/L Ag/Fe/MB dose,
352 CLX initial concentration of 20 mg/L, and a pH of 6.15 [75].
353
354 **2.3 Characterization of Surface Chemistry and Other Properties**
355 For comprehensive understanding of structural features, physicochemical properties,
356 contaminant adsorption mechanisms of biochar-MNPs, it is crucial to apply a series of analytical
357 techniques for extensive characterization as outlined in **Figure 3** and discussed below.
358 FTIR is a powerful tool for the detailed characterization of pristine and nano-modified
359 biochar biochar, providing insights into their chemical structure, functional groups, and
360 modifications, which are vital for their applications in environmental remediation and other fields.
361 For instance, Zhang et al. employed FTIR to reveal numerous surface functional groups of
362 biochar/Au composite material [62]. The spectral region between 3500–3100 cm^{-1} was assigned
363 to the stretching vibrations of O-H (hydrogen bond). Vibrations of C-H stretching (hydrogen bond)
364 were represented in the 3930–2910 cm^{-1} range. The spectral band from 1740–1655 cm^{-1} arose due

to the stretching vibrations of C=O. The 1280–1210 cm^{-1} region corresponded to the stretching of C-O and the presence of aromatic ring. Lastly, the 1000–650 cm^{-1} range is associated with vibrations of M-O, where M denotes metal or Si. Research on studying functional groups on nano-modified biochars is limited. Biochars are known to have numerous functional groups on their surfaces and display heterogeneous compositions [69]. The Surface O/C ratios can undergo substantial modifications, especially at synthesis temperatures between 300-400 °C, impacting the overall surface behavior, and reactivity [76]. The synthesis temperature is pivotal in determining the surface characteristics of both bulk and nano-modified biochars. Biochar generated at lower temperatures (below 400 °C) experiences less severe carbonization and has a lower degree of aromaticity but more surface functional groups, exhibiting enhanced adsorption capabilities. In contrast, biochars created at elevated temperatures (higher than 700 °C) tend to have reduced functional groups due to the removal of heteroatoms like O, N, and S during the charring process [77]. This leads to the presence of polyaromatic rings (characterized by X-ray adsorption spectroscopy) with various functional groups such as O-, S-, and N-functional entities. The hydrophilic properties and negative charges are predominantly attributed to the O-functional groups such as –OH, –COOH, –C=O, and –C–O, as the deprotonation of these groups results in net negative surfaces [78]. Additionally, nano-modified biochars may incorporate N-H bonds and SiO_2. Comparatively, nano-modified biochars feature both polar and non-polar sites along with a plethora of hydroxyl, carboxylic, and carbonyl functional groups.

BET method is utilized to ascertain specific surface area and distribution of pore sizes. For example, Zhang et al. used BET surface analyzer to measure the pore size distribution and surface area of biochar-MnO_2 material [56]. Biochar initially had a minimal BET surface area of 3.5 m^2/g due to its substantial particle size. However, a notable enhancement in its BET surface area was

observed upon the deposition of MnO_2. Biochar loaded with varying amounts of MnO_2 displayed superior specific surface areas compared to unmodified biochar, indicating the augmentation in biochar's specific surface area viacpost-MnO_2 modification. The specific surface area of biochar-MnO_2 escalated from 135.9 m^2/g to 181.5 m^2/g as the MnO_2 loading amount rose from 12.3% to 18.4%. Conversely, a further increase in MnO_2 loading amount resulted in a decline in the specific surface area of biochar-MnO_2. The initial enhancement can be attributed to the incorporation of MnO_2 nanoparticles and the formation of additional micropores. The subsequent reduction at elevated loading is presumably due to the excessive deposition of MnO_2 nanoparticles, potentially causing blockage of pores and disruption of certain micropore structures.

Transmission Electron Microscopy (TEM) and SEM are employed to examine the microstructure, morphology, and the dispersion of metal nanoparticles within biochar-MNPs. For instance, Yang et al. explored the structure and morphology of the biochar/Fe-Mn-S composite utilizing SEM and TEM [79]. SEM imagery revealed that the ternary needle-like Fe-Mn-S mixtures were uniformly distributed on the hierarchical biochar, which possessed numerous regular and vertical channels. TEM images displayed the contours of nanorods, having diameters of several nanometers and lengths varying from several to hundreds of nanometers. Energy-Dispersive X-ray Spectroscopy (EDS) is a technique used for characterizing the elemental composition of materials. If biochar is modified with metal nanoparticles, EDS can confirm the presence of metal elements on the biochar surface and provide information about their concentration and distribution. Wang group adopted EDS to analyze the types and molar fraction of elements in a biochar sample decorated with gold nanoparticles [80]. The distributions of C, O, N, and Au appear to be uniform across the entire surface. Therefore, the combined results from their SEM and EDS analyses effectively demonstrate the embellishment of biochar with gold

411 nanoparticles. In addition, elemental analyzer was also used for analyzing C, H, O, and N. Biochar
412 is predominantly composed of those elements, along with varying concentrations of vital mineral
413 nutrients, and exhibit significant diversity in trace elements. The nitrogen component of biochar is
414 influenced by feedstock and pyrolysis temperature as peptide bonds transform into N-
415 heteroaromatic carbon compounds [69]. Biochar derived from plants usually possesses a richer
416 mineral content.

417 XRD is a versatile tool that can provide a wealth of information about the structural and
418 compositional characteristics of pristine and nano-modified biochars . It can provide valuable
419 insights into the crystalline nature, mineral content, and other structural attributes of the char by
420 analyzing the diffraction patterns. For instance, Wang et al. employed XRD to examine the
421 crystalline phase and structural characteristics of magnetic nano-modified biochar [81]. The XRD
422 analysis of the magnetic nano-modified biochar revealed peaks at 2θ values of 30.2°, 35.6°, 43.3°,
423 53.7°, 57.2°, and 62.9°, which correspond to the γ-Fe_2O_3 crystalline phases. Given the similar
424 diffraction peaks of γ-Fe_2O_3 and Fe_3O_4, the XRD results confirmed the formation of both Fe_3O_4
425 particles within the composite.

426 X-ray Photoelectron Spectroscopy (XPS) is a surface-sensitive analytical technique that
427 provides information about the elemental composition, chemical state, and electronic state of the
428 elements present in the pristine and nano-modified biochars . In 2019, Yang group utilized XPS
429 to examine the surface composition and chemical states of a straw biochar functionalized with
430 lanthanum ferrite ($LaFeO_3$) nanoparticles [82]. The observed peaks at 835.0, 836.2, and 839.7 eV
431 for $LaFeO_3$/biochar correspond to the binding energies of La (3d), suggesting that La is
432 predominantly in its oxidized state. For $LaFeO_3$/biochar, the Fe (2p) peaks are located at 711.5 and
433 725.1 eV, with no discernible shoulder peaks, signifying the Fe ions are in the Fe^{3+} oxidation state.

The 529.8 eV peak is attributed to lattice oxygen atoms, specifically La–O and Fe–O in the $LaFeO_3$ crystal structure. Regarding C in $LaFeO_3$/biochar, the C (1s) signal deconvolution revealed the presence of C–C bonds at binding energies of 283.7, 284.7, and 285.1 eV. This data confirms the successful incorporation of $LaFeO_3$ onto the biochar surface.

Thermogravimetric Analysis (TGA) is a thermal analysis technique used to measure the weight change of a sample as a function of temperature or time under a controlled atmosphere. For example, Zhang et al. employed TGA to study the development of MgO-biochar [83]. The TGA curve for sugar beet tailings revealed four distinct weight loss phases. The first phase, from 30 to 200 °C, was due to residual water molecule evaporation. The subsequent phases, from 200 to 300 °C, 300 to 400 °C, and 400 to 550 °C, were attributed to the decomposition of lignin, cellulose, and hemicellulose, respectively. The TGA curve for $MgCl_2$-treated sugar beet tailings also showed four stages. The initial weight loss was due to the water molecule evaporation from 30 to 100 °C. The subsequent decrease in weight was linked to the breakdown of $MgCl_2$ hydrates from 100 to 250 °C. The third drop in weight was likely due to the formation of magnesium hydroxychloride and the further degradation of hydrates from 250 to 450 °C. The final weight reduction was associated with magnesium oxide formation alongside decomposition of magnesium hydroxychloride from 450 to 650 °C.

VSM provides a comprehensive analysis of the magnetic properties of pristine and nano-modified biochars. For example, Li et al. utilized a vibrating sample magnetometer to assess the magnetic saturation of biochar fabricated with magnesium oxide [84]. The biochar impregnated with magnesium displayed a saturation magnetization of 35.35 emu/g, suggesting it could be readily magnetized. Impressively, even after being submerged in water for five months, this

magnesium-impregnated magnetic biochar could be effectively recaptured using a permanent magnet.

In recent years, innovative methods have been employed to analyze multifunctional composites. Techniques like X-ray Absorption Near-Edge Structure (XANES) and extended X-ray Absorption Fine Structure (EXAFS) are used to provide deep insights into the interaction between the biochar matrix and the nanoparticles, the chemical state of the metals, and the nature of the metal-containing species. Confocal Micro X-ray Fluorescence (μ-XRF) is applied to ascertain the elemental composition and distribution in biochar-MNPs [85]. In a study by Feng et al. they utilized XANES, EXAFS, and μ-XRF to study the characteristics of Fe-biochar [86]. The results from Fe XANES and EXAFS revealed γ-Fe_2O_3 as the dominant Fe species in the composite, with Fe° being detected at temperatures of 600 or 900 °C. The μ-XRF analysis highlighted the dispersion of Fe, S, and Cl within the porous structure, with Fe accumulating on the biochar's surface.

3. Multifunctional Applications of Nano-modified Biochar in Environmental Remediation

Recent studies have delved into nano-modified biochars' diverse applications, such as catalyst, alternative for carbon black, biomolecule carrier, and the removal of various contaminants like agrochemicals and pharmaceuticals [73, 87-89]. Due to its superior adsorptive properties, nano-modified biochar showcases an impressive ability to adsorb a broad spectrum of contaminants. Moreover, nano-modified biochar can also act as a catalyst, promoting the decomposition of organic compounds [90].

3.1 Heavy Metal Removal

Heavy metal contamination poses significant environmental threats, impacting both ecosystems and public health. Nano-modified biochar have gained prominence due to its notable reactivity, high surface area, and ion-exchange capabilities, making it effective in addressing heavy metals like As, Hg, Cr, Cu, Cd, and Pb [91]. Some of these heavy metals are non-essential and known to be detrimental to human health due to their acute and chronic toxic effects. Meanwhile, others are trace elements vital for human well-being but can become harmful if their levels surpass the recommended thresholds [92]. Therefore, effective heavy metal removal strategies including stabilization, microbial remediation, and electrochemical methods, are required in agricultural and environmental research. Several factors, including dosage, presence of other ions, initial heavy metal concentration, and pH, influence nano-modified biochar's performance. The removal processes involve multiple mechanisms, including adsorption, reduction, precipitation, co-precipitation, ion exchange, and oxidation. To elucidate these mechanisms, various analytical techniques as shown in **Figure 3** are employed. The ability of nano-modified biochars to extract heavy metals, the influence of different factors on this removal efficiency, and the underlying mechanisms are comprehensively discussed below.

Numerous studies have demonstrated the effectiveness of nano-modified biochars in removing heavy metals and have assessed its capacity for such removal. To gauge the efficiency of nano-modified biochars in extracting heavy metals, various models such as Langmuir, Freundlich, Langmuir-Freundlich, Temkin, and Sips are used to interpret the collected data. Moreover, models like the first-order kinetic, Elovich, intraparticle diffusion, and second-order kinetic are applied to ascertain the rate of heavy metal removal. In a study by Wan et al., a magnetic biochar, which combines the advantages of both biochar and iron oxides, effectively removed multiple metal(loid) contaminants from the soil [93]. There was a notable reduction in the soil

concentrations of As, Cd, and Pb by 28%, 25%, and 32%, respectively, within 24 h. The Fe_3O_4-modified biochar not only removed the soluble and exchangeable fractions of metal(loid)s but also directly adsorbed the solid fractions. A novel porous biochar activated with K_2CO_3 and infused with nano-zero-valent iron and nano-α-hydroxy-iron oxide was tested for Cd (II) removal from aqueous solution, as reported by Zhu et al [94]. The Cd (II) adsorption onto this composite was best explained by the pseudo second-order kinetic model, suggesting that chemisorption primarily controlled the process. The composite's adsorption rates for Cd (II) were notably rapid. Furthermore, the Langmuir model provided the most accurate fit for the experimental data. The composites' maximum adsorption capacities for Cd (II) were measured at 26.43 mg/g and 22.37 mg/g. When compared to other materials, these composites demonstrated superior efficacy in removing Cd (II). Nano-modified biochar prepared by ball-milling process was used to remove Ni (II) (a model contaminant).[73] The results showed that this ball-milled nano-modified biochar was more efficient in Ni (II) removal than the pristine biochar. It also demonstrated rapid adsorption rates and outstanding Ni (II) absorption capacity, surpassing many commercial sorbents (e.g., Activated carbon, Na-montmorillonite, natural bentonite, and Chlorella sorokiniana). Compared to the pristine biochar, the ball-milled nano-modified biochar exhibited enhanced physical and chemical features, such as increased surface areas and more acidic functional groups. As a result, its ability to remove contaminants surpassed that of the pristine biochar.

The effectiveness and stability of heavy metal removal using nano-modified biochars depend on various environmental factors (soil minerals, feedstock, pyrolysis temperature, nanoparticle content, biochar loading, additives, etc.). For instance, when using biochar loaded with polymetallic nanoparticles, the heavy metal immobilization varies based on the nanoparticle to biochar ratio. In a study reported by Zhang et al., they examined how the addition of iron,

manganese, and cerium oxide (FMCBC) to biochar affected arsenic availability in arsenic-polluted paddy soil [95]. The FMCBC with different mass ratios of 24:2:3:4 ($FMCBC_1$), 24:2:3:8 ($FMCBC_2$), and 24:2:3:10 ($FMCBC_3$) for biochar, Fe, Mn, and Ce respectively, is used for the polluted soil treatment. The findings indicate that FMCBCs enhance the soil's pH and redox potential, leading to a reduction in the bioavailable arsenic forms. Among all treatments, $FMCBC_3$ had the highest reduction in soil-bound arsenic concentration. The conditions under which pyrolysis is conducted, including temperature and gas flow, influence the functional groups and surface area of biochar composites. For example, greater immobilization of Cr is observed at higher pyrolysis temperatures, specifically at 700 °, as noted by Rafique et al [96]. As pyrolysis temperatures rise, groups like alcoholic, aliphatic, and ester on the biochar surface diminish, as observed by Shakya et al [97]. Additionally, hemicellulose breaks down at temperatures above 350 °C, leading to a rapid release of volatile compounds that form pores [98]. As a result, the capacity to immobilize increases with temperature due to enhanced surface area.

3.2 Organic Pollutant Removal

Nano-modified biochars also demonstrated a strong adsorption to organic contaminants. Nano-modified biochar serves as an adsorbent, reductant, and catalyst in removing organic contaminants such as persistent organic pollutants, cancer-causing compounds, pesticides, and antibiotics. This is applied to environmental and agricultural settings because of its strong adsorptive, catalytic, and reductive properties. Recent research indicates that carcinogenic persistent organic pollutants (POPs), such as polycyclic aromatic hydrocarbons (PAHs) and polybrominated diphenyl ethers (PBDEs), can make their way to the food chain. Due to their carcinogenic properties (e.g., low solubility, high molecular weight, and persistence), they pose potential threats to both human health and the environment [99, 100]. Therefore, nano-modified

biochars can serve as an effective tool for removing these pollutants from the environment. For instance, using the liquid phase reduction method, biochar-supported nano zero-valent iron (biochar-nZVI) was synthesized to activate persulfate (PS) for decabromodiphenyl ether (BDE209) removal from soil [101]. The data indicated that at a PS/biochar-nZVI molar ratio of 3:1, pH 3, and 40 °C, 82.06% reduction of BDE209 was achieved in 4 h. In another study, biochar nanocomposites modified with copper oxide (CuO) were synthesized for removing anionic contaminants (reactive red RR120) from water [102]. The work highlighted the robust interaction between CuO and biochar. The embedded CuO nanoparticles in the composite demonstrated superior removal of reactive red due to strong electrostatic attraction. The CuO/biochar nanocomposite showcased the best RR120 removal efficiency at 46%, significantly outperforming the original biochar at 20%. Furthermore, the adsorption process was quick, with an equilibrium time of less than 3 h. The impact of different particle sizes (sub-millimeter, micron-scale, and nano-scale) of biochar derived from corn straw and rice husk was studied on the adsorption of diethyl phthalate (DEP) [103]. The findings revealed that the smaller the biochar particle size, the greater its capacity to adsorb DEP, with nano-scale biochar demonstrating the highest adsorption capability. This increased adsorption in nano-scale biochar was due to its advanced pore structure and larger specific surface area. This suggests that nano-scale biochar's DEP adsorption primarily resulted from pore-filling, rather than interactions like π-π EDA and H bonding observed in larger-sized biochar. Moreover, the nano-scale biochar's ability to adsorb DEP significantly dropped when the initial pH value was reduced from 9 to 3. This was because an acidic environment diminished the surface charge on the nano-scale biochar, causing easier agglomeration of the particles. Phthalate esters (PAEs) are commonly used as plasticizers in plastic products worldwide, leading to the potential bioaccumulation of benzene carboxylic groups. It is crucial to remove

PAEs from environments like soil and sediment. Two prominent plasticizers, Dibutyl phthalate (DBP) and di-(2-Ethylhexyl) phthalate (DEHP), are prevalent residues in soil and are listed as priority pollutants by the US Environmental Protection Agency. In a study, the impact of using the Fe-Mn oxide-modified biochar composite (FM-biochar) was investigated on the quality of wheat cultivated in brown soil contaminated with DBP and DEHP [104]. Introducing FM-biochar notably reduced the absorption of DBP and DEHP by wheat grains, leading to enhancements in the grain quality, specifically in starch, protein, and amino acid levels. Moreover, the results from FM-biochar applications surpassed those from biochar treatments, suggesting that FM-biochar is a potent method to lower the bioavailability of PAEs to wheat grains. In their report Bentley et al. showed improved adsorption capacities for certain organic micropollutants (OMPs) using pine-derived biochars pre-treated with NaOH (pH 9 and 11) and various alkali and alkaline earth metals (AAEMs) like Na, K, Ca, and Mg [105]. This treatment notably increased the surface area of the biochar's micropores, with the most significant improvement observed at pH 11, where the OMP adsorption rate nearly matched that of commercial activated carbon, surpassing untreated biochars. This underscores the effectiveness of specific pre-treatments in enhancing biochar for water treatment applications.

3.3 Nano-modified Biochars' Role in Soil Health and Air Quality

Biochars, as soil amendments, have a significant impact on soil microbial activity, playing a crucial role in enhancing soil health and fertility. Their porous structure and high surface area provide a conducive habitat for diverse microbial communities, facilitating the proliferation of beneficial microorganisms. Biochars are known to influence soil pH, moisture retention, and nutrient availability, all of which are critical factors affecting microbial growth and activity. Notably, they can improve the soil's carbon content, offering a stable carbon source for soil

592 microbes, thus stimulating their metabolic processes. The interaction between biochar and soil
593 microbes often leads to increased soil enzyme activities, indicative of enhanced nutrient cycling
594 and organic matter decomposition. Furthermore, the presence of biochar can help in mitigating the
595 negative effects of soil contaminants, thereby reducing the stress on microbial communities. This,
596 in turn, supports the restoration of microbial diversity and function in degraded soils.

597 Soil microorganisms play a pivotal role in nutrient cycling, breaking down soil organic
598 matter to release nutrients in accessible forms for plants and other microbes, integral to the soil-
599 plant biogeochemical cycle. Any changes affecting these microbes or nutrient cycling can impact
600 ecosystem functionality. Previous findings indicate that nano-modified biochars can transform the
601 composition of soil microbial communities and influence their metabolic capacity, particularly in
602 nutrient cycles like carbon and nitrogen cycling. The presence of nano-modified biochars affects
603 enzyme activities, functional genes, and microbial species involved in nitrogen and carbon cycling.
604 Factors such as soil pH, bioavailable metals, and soil electric conductivity (EC) also impact
605 bacterial communities, hindering phototroph growth and leading to a decrease in carbon fixation
606 genes [106]. Additionally, dissolved organic carbon (DOC) released from biochar forms reactive
607 oxygen species, which can damage cellular components of phototrophs, reducing the relative
608 abundance of carbon fixation genes [107]. The promotion of carbon-degrading organisms by
609 MgO-biochar might be associated with increased cellulase activity, releasing more glucose and
610 thereby boosting the abundance of carbon-utilizing bacteria [108]. In some cases, the rise in certain
611 key orders related to carbon-cycling bacteria in modified biochar treatments could regulate the
612 carbon cycle. In general, the use of nano-modified biochars can influence crop growth and yield,
613 affect the behavior and survival of soil fauna, and alter soil properties like pH and organic matter

[109]. These changes impact the structure of soil microbial communities, which are crucial for processes like carbon and nitrogen cycling.

Nano-modified biochars have also shown significant potential in improving air quality by reducing emissions of harmful gases such as nitrous oxide (N_2O) and methane (CH_4), and sequestering greenhouse gases like carbon dioxide (CO_2) [110-112]. The high surface area and reactivity of nano-modified biochar enable it to adsorb pollutants more efficiently than conventional biochar. By altering the soil's physicochemical properties, nano-modified biochar enhances its ability to retain and degrade airborne pollutants and provides a more favorable environment for microbial communities that break down these gases. This, in turn, contributes to both improved soil health and air quality by decreasing greenhouse gas emissions through the modification of soil microbial activities that influence the nitrogen and carbon cycles [34].

4. Mechanisms of Pollutant Removal by Nano-modified Biochars

As discussed above, nano-modified biochar has been identified as a capable adsorbent for removing specific pollutants, including heavy metals and organic pollutants. Understanding the interaction mechanism between pollutants and nano-modified biochar is vital for linking remediation to the engineering design of these materials [113, 114]. Factors such as pH and material dosage significantly influence the efficiency of dissipation. Extensive research has been carried out to explore the remediation abilities of nano-modified biochar for heavy metals and organic compounds in polluted environments. The proposed mechanisms behind pollutant adsorption are illustrated in **Figure 4** and discussed below**.**

4.1 Mechanisms Involved in Metal Removal

Electrostatic Interactions. Electrostatic adsorption, which is fundamental to the formation of ionic bonds, can be affected by factors such as point of zero charge (PZC), the pH level, and the

heavy metal's ionic and valence radii. In general, elements like chromium, arsenic, and antimony form anionic species in water, and because they are negatively charged, they tend to repel away from the similarly charged surface of unmodified biochar. This electrostatic repulsion limits the capacity of pristine biochar to adsorb these elements. To overcome this limitation, modifications to biochar are often employed to improve its adsorption capabilities. For example, Zhang et al. observed that by loading chitosan into bamboo biochar, the adsorption capacity for hexavalent chromium (Cr(VI)) was significantly enhanced, reaching up to 127 mg/g of biochar. This increase in adsorption was attributed to the electrostatic attraction between the positively charged hydroxyl and amine groups of chitosan and the CrO_4^{2-} ions [115]. Likewise, it was reported that the KOH-modified, nitrogen-enriched biochar (KNB) derived from waste chicken feathers, was effectively utilized for the adsorption of Cd and Pb from wastewater [116]. KOH modification significantly enhanced the adsorption capacities of the biochar. Electrostatic interactions were one of the primary mechanisms for Cd and Pb adsorption by KNB. The modification with KOH enhanced the biochar's surface functional groups, which in turn increased its electronegativity. This enhancement led to a stronger electrostatic attraction to Cd and Pb ions.

The adsorption efficiency of engineered biochar for positively charged heavy metals, such as cadmium, lead, copper, zinc, and nickel, is significantly affected by several factors. These factors include the surface area of biochar, its zeta potential, the nature of minerals present in biochar, the types of functional groups on the biochar surface, as well as the conditions present in the solution. For example, FTIR analysis indicated that Cd^{2+} ions were likely adsorbed onto the surface of iron oxide through electrostatic attraction. Furthermore, the analysis suggested that surface complexation occurs between the Cd^{2+} ions and the Fe_3O_4, involving π-orbital bonding of the metal ions to the π-electrons of the carbon atoms on the surface [117].

Complexation. The surfaces of biochar are enriched with oxygen-containing functional groups such as -OH, -COH, and -COOH. These groups participate in complex formation with metal ions, resulting in the creation of stable metal complexes. For instance, a biochar supported nano-scale zero-valent iron (biochar-CMC-nZVI), which was stabilized with carboxymethyl cellulose (CMC), was utilized for the effective removal of hexavalent chromium (Cr (VI)) from water [118]. The XPS pattern indicated that alcohol, phenolic, aliphatic, and aromatic carbon were present, suggesting that surface complexation played a crucial role in the adsorption process of Cr(VI). The FTIR spectra of the composite post-reaction revealed alterations in the Fe-O, C-O, and O-C=O groups, indicating that complexation and electrostatic attraction occurred between Cr (VI) and the composite's iron-containing functional groups. Also, various materials such as metal oxides, sulfides, and carbon-based nanomaterials have been utilized to enhance biochar for the purpose of extracting Sb from water [119, 120]. In a specific instance, Wang et al. developed biochar infused with lanthanum (La) that possessed magnetic properties [121]. This modification significantly boosted its capacity to adsorb Sb(V), with the adsorption capacity rising from 2.2 to 18.9 mg/g at a neutral pH of 7. Characterization using FTIR and XPS techniques indicated that the primary mechanism behind the improved adsorption of Sb(V) was the formation of a stable inner-sphere La–O–Sb complex within the material's structure.

Ion Exchange. Ion exchange involved the selective exchange of exchangeable metal ions like K^+, Mg^{2+}, Na^+, and Ca^{2+} on the biochar surface with other metal ions. This process is largely influenced by the chemical characteristics of the biochar's surface [122]. The investigation into the mechanism behind Cu (II) removal by Fe_3O_4-alginate modified biochar involved characterizing the composite post Cu (II) adsorption using FTIR and XPS techniques [123]. Shifts and changes in the FTIR spectral adsorption peaks after Cu (II) adsorption suggested interactions between Cu

683 (II) and the composite's functional groups. Moreover, notable shifts in the Cu 2p and Fe 2p peaks
684 in the XPS spectrum regarding post Cu (II) adsorption indicated that ion exchange took place
685 between Cu (II), the composite material, and Fe_3O_4.

686 ***Redox Reactions.*** Some biochars may facilitate redox reactions due to the presence of
687 electron-donating or accepting sites. This can transform the metal ions into a less mobile and less
688 toxic state [124]. For example, an iron-enhanced biochar, was utilized to assess the absorption
689 capacity of As, Cd, and Pb by rice crops. Heavy metals, particularly arsenic in the forms of As
690 (III) and As (V), undergo redox reactions on the surface of biochar due to the strong oxidizing and
691 reducing agents present, resulting in the formation of stable inner-sphere complexation of As on
692 the biochar surface [125].

693 ***Precipitation.*** Biochar can co-precipitate with heavy metal cations, forming insoluble
694 carbonates and phosphates that serve to immobilize heavy metals in soils. Liang et al. enhanced
695 biochar derived from glucose by activating it with KOH and incorporating nitrogen at a pyrolysis
696 temperature of 800 °C [126]. This modification endowed the biochar with a substantial Cr (VI)
697 adsorption capability of 402.9 mg/g. The adsorption process involved a combination of
698 mechanisms, including physical adsorption, electrostatic attraction, surface complexation, and the
699 reduction of Cr (VI) to Cr (III), followed by the adsorption and precipitation of the reduced Cr (III)
700 species.

701 Other mechanisms like physical adsorption and micropore filling can also be involved in
702 heavy metal removal. The porous nature of biochar provides a large surface area that facilitates
703 the physical adsorption of metal ions. This is a non-specific process where metal ions are retained
704 on the surface of biochar through van der Waals forces [122]. In the micropore filling, heavy metal

ions may get trapped within the micropores of the biochar, especially if the pore size is compatible with the size of the metal ions.

4.2 Mechanisms involved in organic pollutant removal

Organic pollutants such as dyes, pesticides, antibiotics, plasticizers, PAHs, and phenols represent significant categories of contaminants in the environment. The primary removal mechanisms for these common organic pollutants by nano-modified biochars are depicted in **Figure 4**. The adsorption of organic contaminants on biochar is controlled by several mechanisms, predominantly including pore-filling, π–π interactions, hydrophobic effects, electrostatic attraction, and hydrogen bonding as discussed below.

Pore Filling. The process of pore-filling plays a crucial role in the adsorption of organic substances onto biochar. The capacity for adsorption is directly proportional to the surface area of the micropores [127]. Chen et al., discovered that the surface area of biochar, which affects the absorption rate of naphthalene (NAP) in solutions, is determined by the temperature at which the biochar is pyrolyzed [128]. Higher pyrolysis temperatures lead to more complete carbonization of the organic components in the biomass, resulting in biochar with a greater degree of carbonization, a larger surface area, and more developed micropores, all of which contribute to a higher sorption rate. Zhu et al. also noted that the extensive surface area and pore volume of carbon-based materials typically enhance the sorption of organic pollutants due to the pore-filling effect [129].

π–π Interactions. Many organic pollutants have aromatic components that can undergo π–π electron donor-acceptor interactions with aromatic regions in biochar. Xie et al. reported a strong correlation between the adsorption of sulfonamides (SAs) and the degree of graphitization of various biochars [130]. They attributed this correlation to the π-π electron donor-acceptor (EDA) interactions that occur between the graphitic surfaces of the biochars (which act as π electron

donors) and the SAs (which act as π electron acceptors), leading to the pronounced adsorption observed. In a different study, a graphene oxide-modified magnetic sludge biochar (GO/$CoFe_2O_4$-S biochar) was developed for the removal of the pesticide imidacloprid [131]. They observed a maximum adsorption capacity of 8.64 mg/g for imidacloprid on GO/$CoFe_2O_4$-S biochar, which they attributed to the increased surface area, the π-π conjugation interactions between the graphitic structures/−OH groups and the pollutant, and the abundance of available binding sites such as C–O, Fe–O, and Co–O.

Hydrophobic Interactions. The non-polar domains of biochar can sequester non-polar organic compounds, removing them from the aqueous phase. For example, the use of engineered biochar for the elimination of antibiotics has been well-documented. Zeng et al. activated coffee ground-derived biochar with H_3PO_4 to target the adsorption of sulfadiazine [132]. The biochar's modification led to a significant uptake capacity for sulfadiazine, reaching 139.2 mg/g, which was primarily due to the hydrophobic effect and π–π electron donor-acceptor (EDA) interactions.

Additionally, hydrogen bonding, electrostatic forces, and hydrophobic interactions are the prevalent processes for the removal of organic pollutants in natural environment. For instance, a sulfur-doped biochar derived from tapioca peel waste proved to be an effective adsorbent for malachite green (30.2 mg/g) and rhodamine B (33.1 mg/g), with electrostatic attraction, surface complexation, and hydrogen bonding identified as the primary mechanisms for their removal [133].

5. Future Directions

The next phase of research should build upon the solid foundation laid by current studies, pushing the boundaries of innovation towards sustainable and scalable solutions. Below are

outlined key areas where focused efforts could yield substantial advancements, further solidifying the role of nano-modified biochar in environmental remediation.

5.1 Advancements in Synthesis Methods

The synthesis of nano-modified biochar is a critical area where innovation can significantly amplify its utility and efficiency. Future research should prioritize the development of novel synthesis methods that enhance the intrinsic properties of nano-modified biochar, such as increased surface area, porosity, and reactivity. These improvements could dramatically elevate its adsorption capacity and effectiveness in pollutant removal. There is a growing need to refine current methods to make them more energy-efficient and environmentally benign. The exploration of low-temperature synthesis processes, or the utilization of renewable energy sources, could contribute to more sustainable production practices. Another key area is the scalability of synthesis methods. The transition from lab-scale to industrial-scale production often presents challenges in maintaining the quality and characteristics of nano-modified biochar. Research into scalable production techniques that can be consistently reproduced will be crucial for widespread application. The field could also benefit from the exploration of new synthesis techniques, such as microwave-assisted pyrolysis or hydrothermal carbonization, which might offer unique advantages in terms of yield, quality, and environmental impact. Tailoring the properties of nano-modified biochar to target specific pollutants or environmental conditions is another promising direction. This might involve manipulating the feedstock or the synthesis conditions to produce nano-modified biochar with specific characteristics suited to target remediation needs.

5.2 Enhancing Environmental Sustainability in Biochar Production

In addressing the environmental concerns associated with the calcination process inherent in biochar production, recent research underscores the importance of implementing innovative

strategies to mitigate emissions and optimize sustainability. Optimization of pyrolysis conditions has been demonstrated to significantly reduce the emission of greenhouse gases and volatile organic compounds [134]. The integration of advanced emission control technologies, such as scrubbers and catalytic converters, further enables the effective capture of pollutants, thereby minimizing atmospheric emissions [135]. Selecting appropriate biomass feedstocks, particularly those that are waste-derived, can also play a crucial role in reducing environmental impacts by diverting waste from landfills and reducing methane emissions [136]. Importantly, life cycle assessments of biochar systems reveal that, when considering the full production and application cycle, biochar can offer a net positive environmental impact, particularly in terms of carbon sequestration and soil health improvement, thereby offsetting the emissions generated during its production [137]. These strategies collectively underscore the multifaceted approach required to address the environmental challenges of biochar production, emphasizing the need for a balanced assessment of both its potential environmental impacts and its substantial benefits in environmental remediation applications.

5.3 Assessing and Managing the Environmental Fate of Pollutants in Nano-modified Biochar Applications

It is pivotal to address the bioavailability and environmental fate of pollutants once immobilized by these advanced materials. While nano-modified biochar demonstrates a high capacity for adsorbing and immobilizing various pollutants due to its large surface area and enhanced reactivity, the potential for re-release of these pollutants into the environment is a subject of ongoing research. Factors such as environmental conditions, the properties of the pollutants, and the physical and chemical characteristics of nano-modified biochar play crucial roles in this process. An illustrative example of this is the significant impact that environmental pH levels have

on the biogeochemical activities of nano-modified biochar in environmental remediation efforts. Research by Gaya et al. demonstrates that in acidic (low pH) environments, the functional groups on nano-modified biochar are prone to protonation, leading to the formation of H^+ ions [138]. This process initiates a competitive interaction between H^+ ions and cationic pollutants for the cation adsorption sites available on the nano-modified biochar , thereby reducing its pollutant adsorption capacity, as further evidenced in the studies by Mahmoud et al. and Park et al [139, 140]. Additionally, it was reported that at high pH levels, the surface of nano-modified biochar , which carries a negative charge, exhibits a lower affinity for negatively charged or neutral pollutants due to electrostatic repulsion, further complicating the dynamics of pollutant adsorption. Therefore, it becomes crucial to monitor pH variations in the environment when employing nano-modified biochar for the purpose of environmental remediation. Adding to this, when nano-modified biochar is introduced into the environment, its ability to stabilize a significant number of pollutants can considerably decrease their transport and bioavailability, thereby mitigating their impact on ecosystems and human health. Yet, the possibility of immobilized pollutants being released back into the environment remains an unresolved question. This underscores the challenge of retrieving used biochar from the environment for remediation purposes. Given the nano-scale size of nano-modified biochar, recycling it poses an even greater challenge. Consequently, the development of nano-modified biochar recovery technologies and a deeper understanding of its environmental risks for post-pollutant adsorption warrant further investigation. In addition to the factors previously mentioned that influence the environmental fate of pollutants immobilized by nano-modified biochar, the interactions between nano-modified biochar and plant systems represent a critical area for future investigation. Plants can play a significant role in the uptake and transport of biochar nanoparticles, potentially leading to bioaccumulation and trophic transfer in the food

chain. Consequently, future studies should assess the extent to which nano-modified biochar is taken up by root systems, its translocation within plant tissues, and its eventual fate in terms of bioaccumulation and potential food chain transfer. Such research is vital to fully understand the ecological risks and to develop strategies to mitigate any negative impacts, ensuring that nano-modified biochar applications in environmental remediation do not inadvertently pose a threat to ecosystem health or food safety.

5.4 Material Modification and Functionalization

The versatility of nano-modified biochar can be significantly enhanced through strategic material modification and functionalization. This area of research holds the key to unlocking the full potential of nano-modified biochar's in environmental applications, particularly pollutant removal. Tailoring the surface properties of nano-modified biochar through chemical or physical modifications can significantly improve its affinity for specific pollutants. Investigating various functional groups that can be introduced onto the biochar surface could lead to more selective and efficient adsorption processes. The development of nano-modified biochar-based hybrid materials is an exciting prospect. By combining nano-modified biochar with other materials such as metal oxides, graphene, or polymers, the resultant composites could exhibit synergistic properties, enhancing both capacity and selectivity for pollutant removal. Future research should focus on customizing nano-modified biochar for specific pollutants, which could involve modifying it to improve its efficiency in adsorbing and breaking down complex organic compounds or heavy metals. Improving the stability and reusability of nano-modified biochar is crucial for its practical application. Studies that investigate how modifications impact the durability and regeneration capacity of nano-modified biochar can lead to more sustainable and economically viable solutions.

6. Conclusion

In this review, we have comprehensively discussed the modification and functionalization of nano-modified biochars, delving into various preparation and characterization techniques of biochar-MNPs. We also shed light on the use of these nano-modified biochars in remediating natural environments where water and soil are contaminated with both heavy metals and organic substances. The review also reveals that certain characteristics of nano-modified biochars, such as enhanced surface area, addition of functional groups, and improved electron transport capacity, are crucial in boosting its efficiency for multifaceted decontamination purposes. Future studies should delve deeper into the biogeochemical behavior of nano-modified biochar, exploring how its interactions with soil and water matrices affect the long-term stability and efficacy of contaminant immobilization. Additionally, understanding the bioavailability of pollutants in the presence of nano-modified biochar is crucial for evaluating its environmental safety and effectiveness. Research in these areas will not only broaden our comprehension of nano-modified biochar's environmental interactions but also refine its applications for sustainable remediation. As nano-modified biochar continues to emerge as a versatile tool for environmental challenges, its integration into broader ecological conservation strategies offers a promising path towards achieving a cleaner and more sustainable planet.

Data availability

No data was generated or used for the research described in the review article.

CRediT authorship contribution statement

Neda Arabzadeh Nosratabad: Writing- Original draft preparation and editing. **Qiangu Yan:** Conceptualization, Writing- Reviewing and Editing. **Zhiyong Cai:** Conceptualization, Writing- Reviewing and Editing. **Caixia Wan:** Conceptualization, Supervision, Writing- Reviewing and Editing.

Declaration of Competing Interest

The authors declare that they have no competing interests.

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1257 **Table 1.** Summary of Biomass Modifications for Metal Nanocomposite Synthesis and Applications

Feedstock	Metal/Organic or Inorganic Precursor	Pyrolysis Condition	Formed Nanocomposite	Application	Adsorption Efficiency	Adsorption Capacity	Ref
Corn stalk	$FeCl_3$ pretreatment in an electric field	600 °C	EC-Fe_3O_4-biochar	Adsorption of Pb(II) ions from wastewater	Not mentioned	113 mg/g	[48]
Wheat straw	Bi_2O_3 and hydrochloric acid	400-600 °C	Bismuth-impregnated biochar	Absorption of P, As(III), and Cr(VI)	Not mentioned	P = 125.40 mg/g As(III) = 16.21 mg/g Cr(VI) = 12.23 mg/g	[49]
Paper and paper sludge (PPS)	$FeCl_3$	750 °C	Fe_2O_3-biochar	Removal of methyl orange (MO) from contaminated wastewater	52.79% higher than that of pristine biochar	20.53 mg/g	[47]
Acacia Nilotica seed shell ash (ANSA)	Nickel and zinc salts	550 °C	$Ni_{0.5}Zn_{0.5}Fe_2O_4$ on ANSA	Adsorption of Pb(II) ions from water	94.8%, using 0.05 g of adsorbent and 50 mg/L Pb (II)	maximum monolayer adsorption capacity of 37.6 mg/g	[50]
Zinc-contaminated corn stover biomass	Zinc compounds	Slow pyrolysis (different temperatures: 500, 600, 700, and 800 °C)	ZnO/ZnS-modified biochar	Absorption of Pb(II), Cu(II), and Cr(VI)	~ 99%	Pb(II) = 135.8 mg/g Cu(II) = 91.2 mg/g Cr(VI) = 24.5 mg/g	[46]

Cow dung	Magnesium	600 °C	MgO-biochar Composite	Reduce leaching of phosphorus from soil	89.25%	30 mg/g	[51]
Orange peel powder (OPP)	Ferric chloride hexahydrate and ferrous sulfate heptahydrate	Coprecipitation (low-temperature method)	Fe_3O_4−OPP (MNP−OPP)	Removal of cadmium ions from aqueous solutions	82%	76.92 mg/g	[52]
Crab shells	Naturally present calcium in crab shells	300 to 900°C	Calcite-based and lime-based CRB	Phosphorus removal or recovery from wastewater	26% to 100% for a phosphate solution 1% to 63% for anaerobic digestion effluent	Not mentioned	[53]
Carboxyl-functionalized CNT-impregnated biochar	Carbon nanotubes (CNT)	Slow pyrolysis (600 °C)	biochar-CNT	Sorption of methylene blue	47%-64%	2.4 mg/g for HC-CNT-1% 5.5 mg/g for biochar-CNT-1%	[54]

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1263 **Table 2.** Overview of Modified Biochar Nanocomposites: Feedstock Sources, Metal Precursors, Pyrolysis Techniques, and Remediation

1264 Applications

Feedstock	Metal/Organic or Inorganic Precursor	Pyrolysis Condition	Formed Nanocomposite	Application	Adsorption Efficiency	Adsorption Capacity	Ref
Invasive water hyacinth	Manganese Dioxide	Slow pyrolysis at low temperature (450 °C)	MnO_2-biochar	Adsorption of Cd(II), Cu(II), Zn(II), Pb(II)	Not mentioned	Cd (II) = 232.5 mg/g Cu (II) = 248.9 mg/g Zn (II) = 239.4 mg/g Pb (II) = 249.2 mg/g	[56]
Sewage sludge	Zero-Valent Iron	600 °C for 3 h, 10 °C/min in N_2 atmosphere	SSB-nZVI	Removal of Cr^{6+} and Pb^{2+} ions	Cr^{6+} = 90% Pb^{2+} = 82%	Not mentioned	[58]
Miscanthus	Amorphous Iron (hydr)oxide	500 °C for an h, 10 °C/min in a N_2 atmosphere	Fe-modified biochar	Arsenite adsorption	Not mentioned	Arsenite powder = 56.06 mg/g Arsenite beads = 47.90 mg/g	[59]
Bamboo	Mg-Al and Mg-Fe LDHs	600°C for 2 h	Mg-Al/Mg-Fe LDH-modified biochar	Phosphate removal from water	95%	Mg-Fe LDH (40%)/biochar= 7.58 mg/g Mg-Al LDH (40%)/biochar= 13.11 mg/g	[60]

Sunflower husk	Silver Nanoparticles (AgNPs)	650 °C	AgNPs-Modified biochar	Tetracycline removal	42.04%	9.55 mg/g	[61]
Tea waste	Gold Nanoparticles (AuNPs)	500 °C for 2 h	biochar/Au Composite	Electrochemical sensor for nitrite detection	~ 100%	Not mentioned	[62]
White myoga ginger (W)	Gold Nanoparticles (AuNPs)	700, 850, and 950 °C for 2 h	W-biochar/Au	Detection of hydroquinone and catechol	~ 95-100%	Not mentioned	[63]
Wetland plant Reed	$FeCl_3$	600 °C for 90 min	nZVIr-biochar	Removal of Pb^{2+}, Cd^{2+}, Cr^{6+}, Cu^{2+}, Ni^{2+}, and Zn^{2+} from polluted water sources	Pb^{2+}, Cu^{2+}, Cd^{2+}, Ni^{2+}, and Zn^{2+} > 98% Cr^{6+} = 48.9%	Pb^{2+} = 38.31 mg/g Cu^{2+} = 30.37 mg/g Cr^{6+} = 23.09 mg/g Cd^{2+} = 39.53 mg/g Ni^{2+} = 47.85 mg/g	[64]
Pineapple peel waste	Fe_2O_3, Lanthanum Hydroxide ($La(OH)_3$)	300 °C	Magnetic biochar with $La(OH)_3$	Phosphate adsorption	> 96.04%	101.16 mg/g	[65]
Bark chips	Chitosan and nanoclay	600 °C at 10 °C/min and 2 h residence time under N_2 flow	Organic-inorganic composite of chitosan, nanoclay, and biochar	Immobilization of Cu, Pb, and Zn	MTCB reduces the metal leaching from the soil by 100, 100, and 52.29% for Cu^{2+}, Zn^{2+}, and Pb^{2+}, respectively	Cu^{2+} = 121.5 mg/g Pb^{2+} = 336 mg/g Zn^{2+} = 134.6 mg/g	[66]

Waste pinus needles (*Pinus roxburrghii*)	$FeCl_3$/g-C_3N_4/$FeVO_4$	600 °C at 10 °C/min	g-C_3N_4/$FeVO_4$/Fe@NH_2-biochar	Removal of methyl paraben (MeP) and 2-chlorophenol (2-CP)	98.4% degradation of MeP 90.7% degradation of 2-CP	Not mentioned	[74]
Sawdust	Ag/Fe Nanoparticles	500 °C	biochar-supported Ag/Fe nanoparticles (Ag/Fe/MB)	Removal of cephalexin from aqueous solution	More than 86% of CLX was removed by Ag/Fe/MB in 90 min	Not mentioned	[75]

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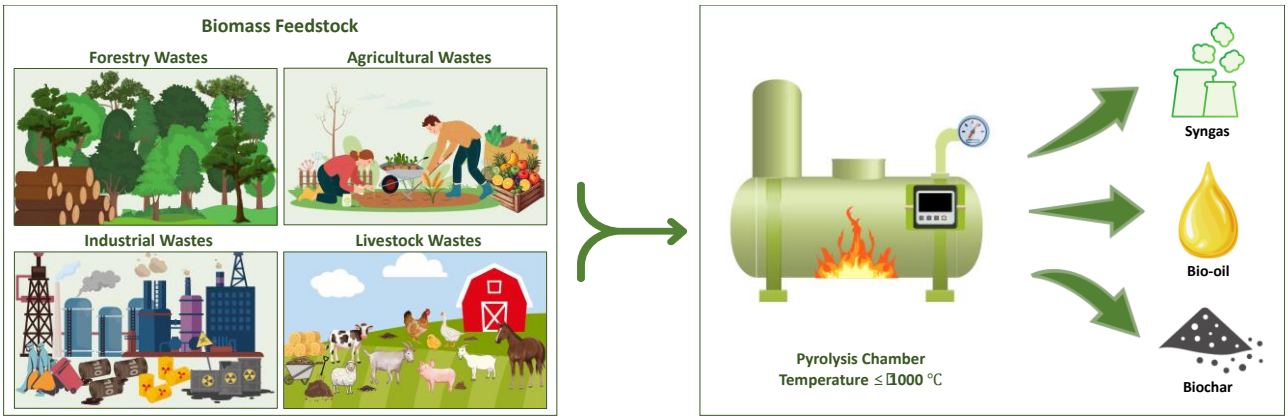
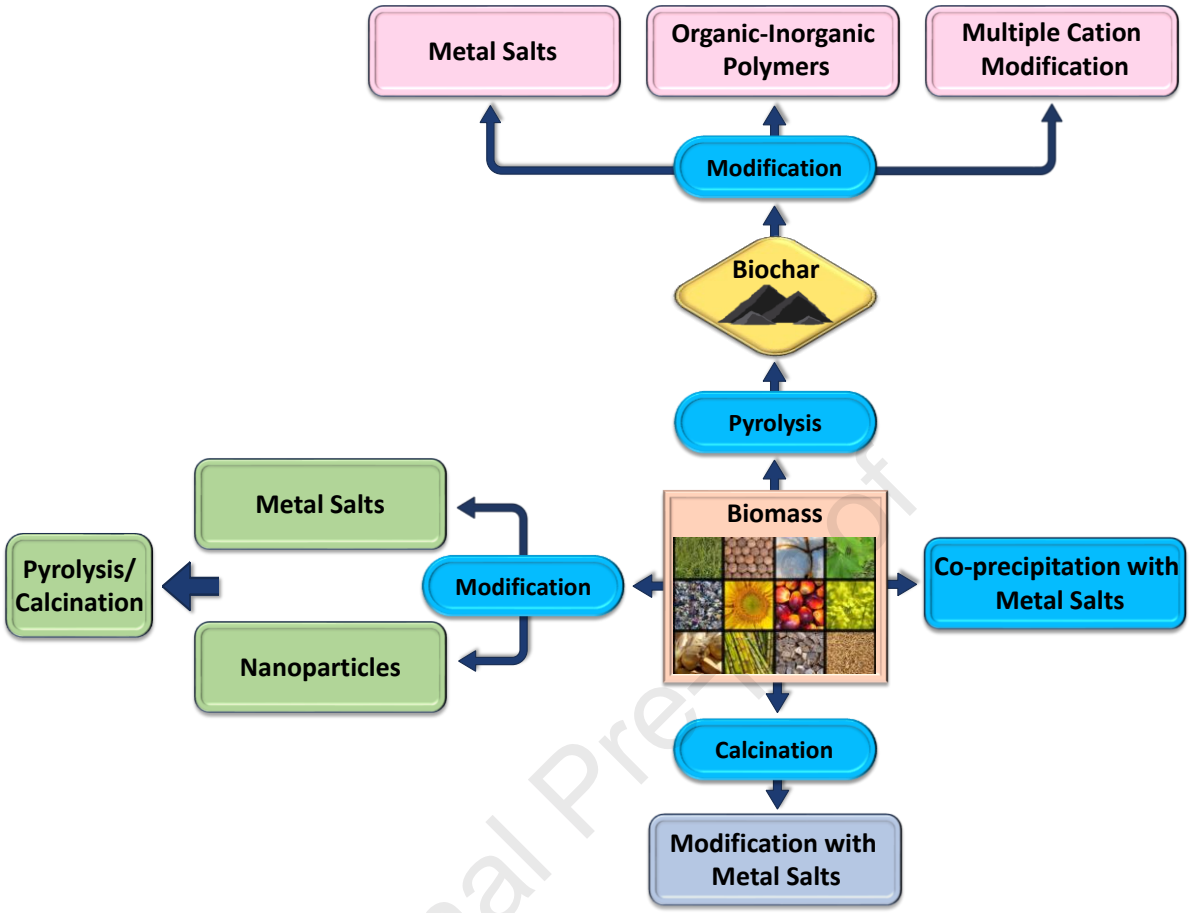


Figure 1. Feedstocks and pyrolytic outputs of biochar.



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1271 **Figure 2.** Schematic diagram of different modification techniques to prepare nano-modified

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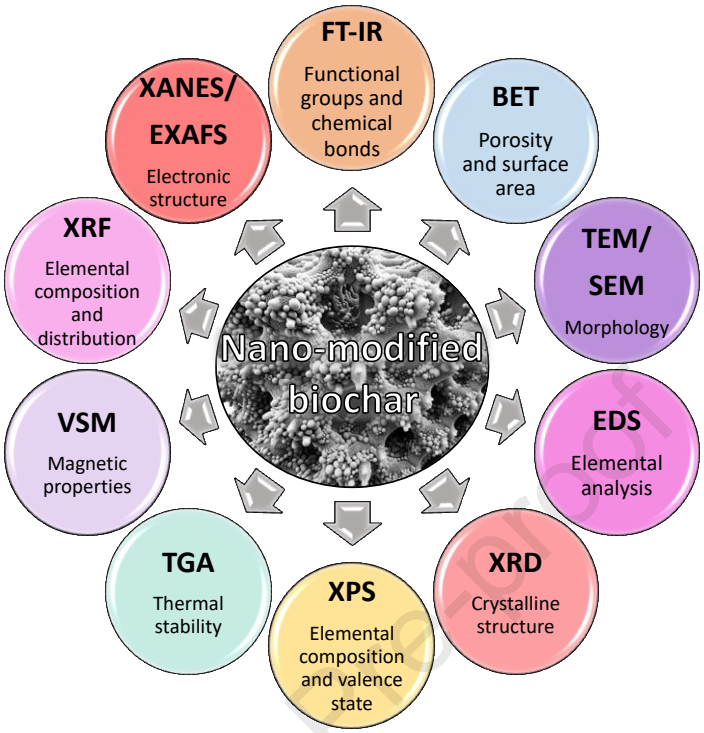

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1275 **Figure 3.** Characterization techniques for nano-modified biochars. Fourier Transform Infrared
1276 Spectrometry (FT-IR), Brunauer–Emmett–Teller (BET) method, Transmission Electron
1277 Microscopy (TEM), Scanning Electron Microscopy (SEM), Energy-Dispersive X-ray
1278 Spectroscopy (EDS), X-ray Diffraction (XRD), X-ray Photoelectron Spectroscopy (XPS),
1279 Thermogravimetric Analysis (TGA), Vibrating Sample Magnetometer (VSM), Confocal Micro X-
1280 ray Fluorescence (μ-XRF), X-ray Absorption Near-Edge Structure (XANES), and extended X-ray
1281 Absorption Fine Structure (EXAFS).

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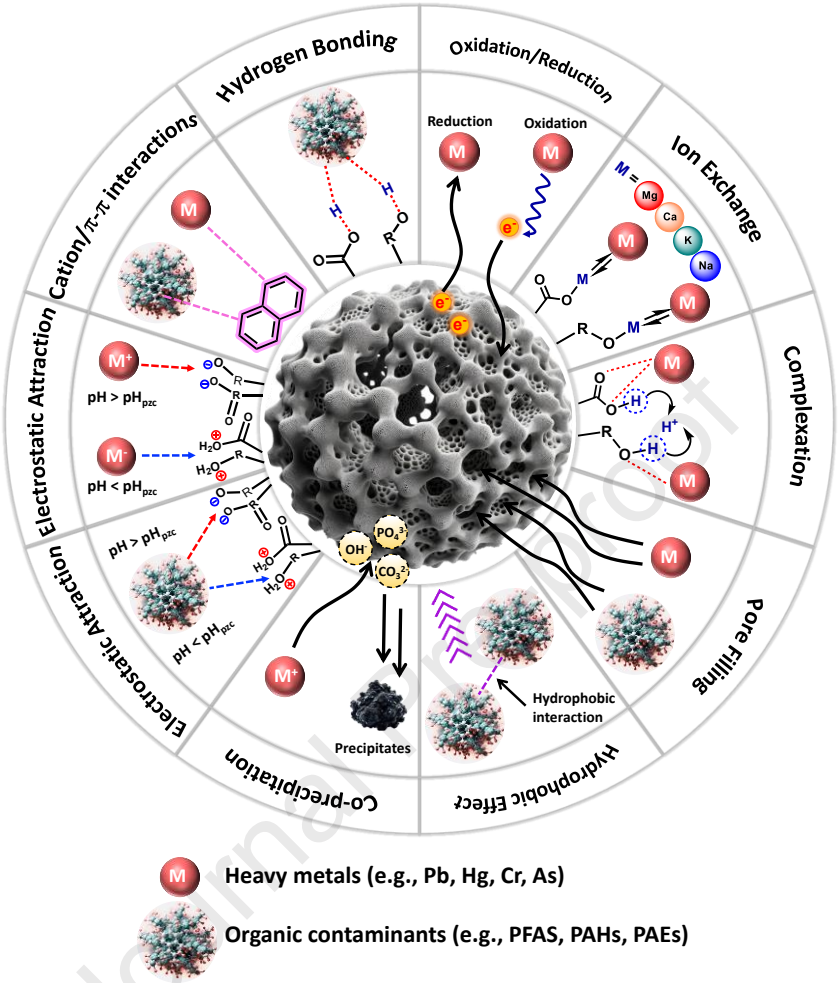

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1284 **Figure 4.** Interaction mechanism between heavy metals and organic pollutants with biochar

1285 particles.

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Highlights

- The review provides a comprehensive summary of the preparation and synthesis techniques for nanomaterial-modified biochars.
- The review also details the characterization methods used for nanomaterial-modifoed biochars, highlighting how these techniques are crucial for understanding their structural and functional attributes.
- The review offers valuable insights into the use of biochar-supported metal nanoparticles (biochar-MNPs) for environmental remediation.
- The review discusses the underlying mechanisms by which nanomaterial-modified biochars effectively remove a wide range of contaminants.

Declaration of interests

☒ The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

☐The authors declare the following financial interests/personal relationships which may be considered as potential competing interests: